



Assessment of Acute Toxic Effects of Tetrabromobisphenol A on Brine Shrimp and *Chlorella* sp.

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Authors' contributions

This work was carried out in collaboration among all authors. Author PMN designed the methodology, performed the investigation, conceptualization, resources, data curation, formal analysis, and wrote the original draft. Author VRK reviewed and edited the manuscript. Author TN contributed to conceptualization, review, editing, and supervision of the study. All authors read and approved the final manuscript.

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ABSTRACT

Tetrabromobisphenol A (TBBPA) is a common brominated flame retardant used in multiple consumer products, such as electronic devices, textile materials, and plastics. Its high utilization makes it widespread in aquatic ecosystems which raising concerns regarding its possible impacts on the environment. Due to its serious threats to human health and the environment, it has become a major concern for the entire population. The objectives of this study were thereby to morphologically investigate the acute toxicity of TBBPA on two aquatic organism brine shrimp and

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Chlorella sp. The results indicated that the addition of TBBPA noticeably inhibited the growth of *Chlorella* sp. with inhibition rates between 40 at the concentration of 6 mg/L. Additionally, TBBPA disrupted the intracellular chloroplast structure at the individual level. Subsequently, *Artemia salina* was treated with TBBPA, which resulting a moderate mortality rate of 58% after 24 hours. The LC₅₀ and LC₉₀ were measured as 128.21 µg/mL (range: 107.86–182.17 µg/mL) and 189.58 µg/mL (range: 150.43–299.35 µg/mL), respectively. This result demonstrates the toxicity of TBBPA against *Artemia salina* and *chlorella* sp. further it can pose a higher risk to aquatic environment.

Keywords: Brine shrimp; TBBPA; emerging pollutant; *Chlorella* sp.

1. INTRODUCTION

Emerging pollutants are synthetic or naturally occurring chemicals that are not commonly monitored in the environment but have the potential to enter into the human body and cause adverse ecological and human health effects (Pereira *et al.*, 2015). The world is facing problems with a wide variety of pollutants and contaminants from various developmental activities. The population explosion in the world has resulted in more widespread water pollution. The concern about the quantity and quality of waste generated and discharged into natural water bodies has recently indicated the need for different strategies to address water quality challenges in the regions (Akhtar *et al.*, 2021). Pollutants enter the environment from different human activities and spread across environmental matrices. Significant progress has been reached in the detection and analysis of trace pollutants in recent decades, due to the ongoing development and refinement of specific techniques. However, there are many different harmful substances that we still haven't found, and we need to identify and measure them in different parts of the environment and in living organisms. These pollutants can be movable and persistent in water, air, soil and sediments, even at low concentrations. Comprehensive data regarding their destiny and behavior in the ecosystem, together with dangers to ecological and human health, remain insufficient. Moreover, the ecotoxicological significance of some emerging micropollutants like Tetrabromobisphenol A (TBBPA) are remains largely unknown, because satisfactory data to determine their risk often do not exist (Gavrilescu *et al.*, 2015).

Pharmaceuticals, pesticides, flame retardant, veterinary and personal care products, nanoparticles (NPs), and nanomaterials (NMs) all fall under the broad category of emerging pollutants. These are commonly derived from municipal, agricultural, and industrial wastewater

sources and pathways (Aravinth *et al.*, 2023). They pose a growing threat to both surface and groundwater quality, and there is an urgent need to better understand their environmental behaviors. Significant research has been performed worldwide in attempts to obtain information regarding their occurrence, fate, and effects on health (Rasheed *et al.*, 2019). Statistics published by EUROSTAT in 2023 revealed that around 68 million tonnes of the manufactured chemicals are environmentally harmful. Over 70% of these are chemicals with significant environmental impact (Duarte *et al.*, 2021; Ntomi *et al.*, 2025). Furthermore, human activities have resulted in contamination of water resources with biological micropollutants, such as viruses and bacteria. These agents, known as emerging or reemerging pathogens, have sparked renewed awareness due to their potential pathogenicity (Ding *et al.*, 2017). Biological micropollutants, such as enteric bacteria, mycoplasmas, viruses, and protozoa, are the source of many waterborne diseases and remain a major cause of death worldwide (Olaolu *et al.*, 2014).

The European Commission (ECB, 2006) and the WHO (1995) have previously deemed exposure to TBBPA to be insignificant. Which is unregulated brominated flame retardant produced in large quantities and used in numerous consumer products. TBBPA exposure is believed to be much below the predicted no-effect thresholds (Colnot *et al.*, 2014). Many studies have revealed that small amounts of TBBPA can disrupt the, thyroid hormones and endocrine system (Yu *et al.*, 2019). The National Toxicology Program (NTP) conducted a recent two-year cancer bioassay of chronic TBBPA treatment in female Wistar Han rats and found a link between TBBPA exposure and an increased frequency of uterine epithelial cancers (NTP, 2014). Based on significant evidence of cancer development in experimental animals, the International Agency for Research on Cancer (IARC) has lately reclassified TBBPA as category

2A, meaning it is "Probably carcinogenic to humans" (IARC, 2018). Studies have thus raised questions about the possible negative effects of TBBPA exposure on people and animals. This study aimed to look at the morphological changes in *Chlorella* sp. and brine shrimp and assess the ecotoxicological impact after being exposed to TBBPA and uncover how it causes harm to primary producers.

2. MATERIALS AND METHODS

2.1 Chemicals and Glassware's

Tetrabromobisphenol A (TBBPA) (>99%) was purchased from Sigma ALDRICH®. Glass wares used for this work, like conical flasks and test tubes (7 mL and 15 mL), were purchased from Glassco®. Centrifugation Falcon tubes (15 mL) and microfuge tubes (1.5 mL and 2 mL) were purchased from TARSONS®. Other glasswares like Petri dishes and conical flasks were purchased from BOROSIL®. All the other chemicals used in this study were analytical grade.

2.2 Sample Collection

Algal samples were collected at midday, a time when photosynthetic activity is considered to be at its highest level during rainy season in sterile glass containers from the Kaveri River in Erode; where much factory wastewater is released.

2.3 Isolation and Identification of *Chlorella* sp.,

The algal samples were gathered in clean bags filled with chu10 medium and then the samples were spun in a centrifuge at 10,000 rpm for 10 minutes to collect the cell pellet. A 100 µL sample was taken with a micropipette and spread over the Chu 10 media in a petri dish aseptically to avoid contamination (Rippka *et al.*, 1979). The petri dishes were then placed in the Germplasm (DMBB laboratory) to grow under a light and dark cycle of 16 hours of light and 8 hours of dark using white fluorescent tubes. The plates were checked after 2-3 days for algae growth, examined under a microscope, and the single pure algal colonies were picked with sterile toothpicks and inoculated into sterile glass test tubes that contained about 1 mL of sterilized Chu's 10 medium. Then again, the inoculated tubes were incubated in germplasm for 3–4 days and checked under a microscope for purity, and

the same culture was transferred to fresh, sterile tubes to maintain its purity. All the isolated pure cultures were photo-documented under an inverted light microscope (Micros-MCX300LED, Austria).

2.4 Artemia Hatching

The Department of Marine Science, Bharathidasan University, provided the *Artemia salina* eggs. The hatching of *Artemia salina* eggs was done by creating the culture medium having seawater with salinity range from 32-35 ppt. Furthermore, weighing *A. salina* cysts as much as 0.5 g and was put into sea water and that functioned as a culture medium for the eggs to hatch (pH 8.4; light 16:8 h; temperature 26 ± 2 °C). Mixing was done carefully to avoid the harm of the cysts. Aerator and the light have been installed to increase the light and dissolved oxygen in the culture container.

2.5 Preliminary Toxicity Assay

A preliminary cytotoxic assay is also called the *Artemia salina* lethality assay. *Artemia salina* cysts were cultured in salt-water under ideal circumstances with constant aeration in order to perform the experiment. Twenty-Five healthy nauplii were moved to separate 24-well plate with varying concentrations of the compound (20, 40, 60, 80 & 100 µg/mL of TBBPA) after they had hatched from the cysts. After 24 h of the incubation, the mortality rate was recorded using abbott's formula (Abbott, 1925), which presented in 1925, to determine the assay's fatality %.

$$\text{Mortality (\%)} = X/Y * 100$$

Y denotes the quantity of deceased larvae post-treatment, while X signifies the count of viable larvae in the control group.

2.6 Acute Toxicity of TBBPA to *Chlorella* sp.

The experiment was conducted using a 250 mL sterile conical flask. The experimental flasks were sterilized, and 90 mL of Chus 10 medium was added to each flask. Once the optical density reached 0.2, 10 mL of *Chlorella* sp. was added to each flask along with varying amounts of TBBPA (2, 4, 6, 8, and 10 mg/L) to test for toxicity at regular intervals (0, 3, 6, 9, 12, and 15 days).

3. RESULTS AND DISCUSSION

3.1 Sample Collection & Identification

Isolation and identification of *Chlorella* sp. from the Kaveri River bed (11.358668, 77.747271) in Erode, Tamil Nadu (Fig. 1). The cells were isolated, non-motile, and spherical, measuring

around 4–6 μm in diameter, and included a conspicuous cup-shaped chloroplast. The cell walls consisted of a singular smooth layer. The strain displayed morphology characteristic to the genus *Chlorella*. Morphological characterization using microscopic images of the pure strain has verified the species of *Chlorella* sp. (Fig. 2).

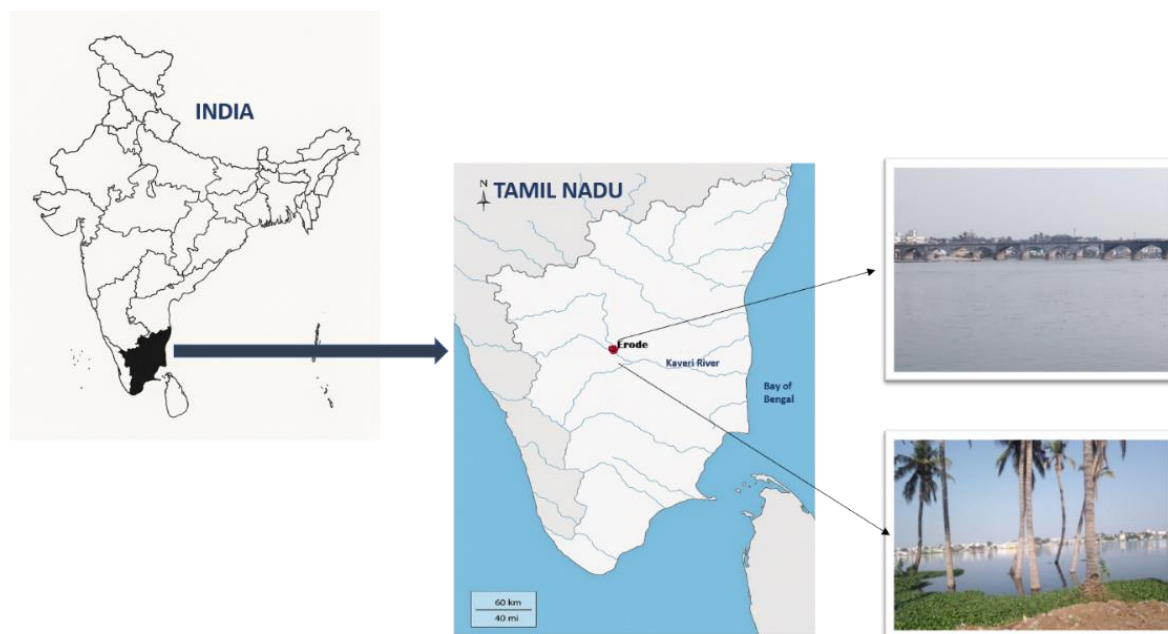


Fig. 1. Sample collection from River Kaveri at Erode

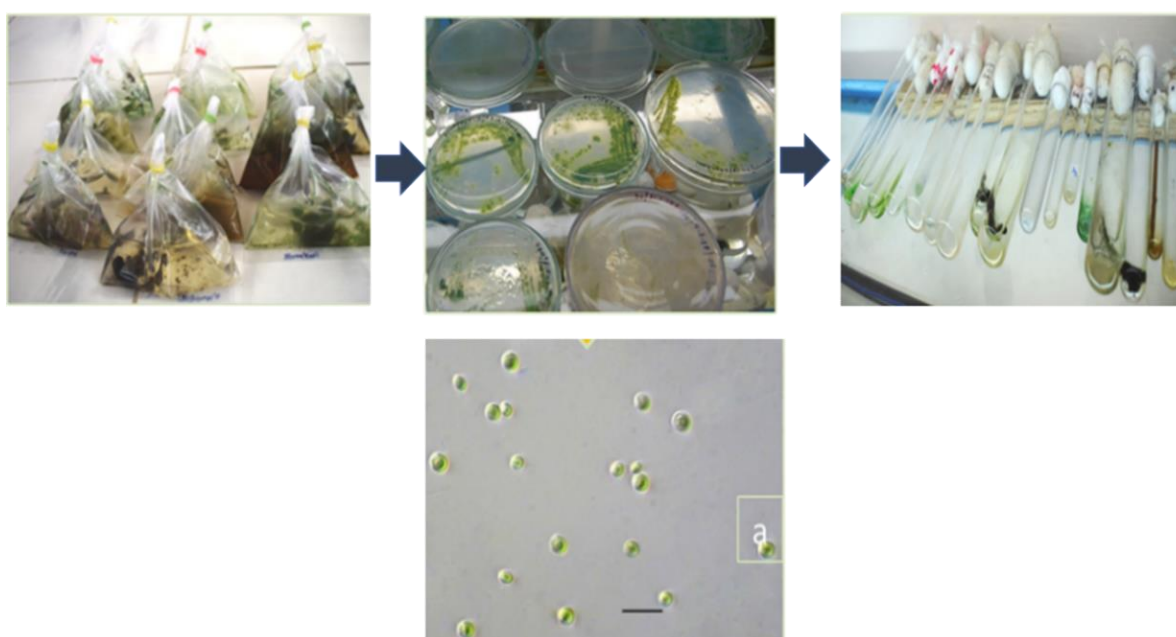


Fig. 2. Process of Isolation and Microscopic images of isolated strains of *Chlorella* sp., (Photomicrograph of 40X magnification)

3.2 Acute Toxicity of TBBPA Pollutant against *Chlorella* sp.

Fig. 3 Shows the concentration of *chlorella* sp. in the treatment groups was considerably lower than that in the control group, whereas algal cells exhibited minimal growth when TBBPA concentrations touched 4 mg/L. As the amount of TBBPA increased, it slowly hindered the growth of *Chlorella* sp., showing that TBBPA negatively influences the proliferation of *Chlorella* sp. The growth of *Chlorella* sp. was better than the control sample when the TBBPA dose was lower at 2 mg/L (Fang et al., 2018). These results are similar to earlier studies that found low amounts of ibuprofen help the growth of the freshwater diatom *Navicula* sp. In the same way, Kurade et al. (2016) observed that between the second and tenth days, there was an increase in the number of cells, indicating a positive interaction between *Chlorella* sp. and DBP at lower concentrations. According to these findings, *chlorella* sp. growth was reduced at higher DBP concentrations (50 and 100 mg/L) but stimulated at low concentrations (5–20 mg/L). The most noticeable indication of physiological damage brought on by pollution is the suppression of microalgal growth (Zhao et al., 2017). The exposure of *Chlorella* sp. to TBBPA concentrations of 8 and 10 mg/L led to significant pigment degradation and subsequent cell mortality, demonstrating severe toxicity at these concentrations.

3.3 Morphological Change in *Chlorella* Sp.

To explore morphological differences between the *chlorella* cells in the control group and those exposed to TBBPA treatments, a light

microscope (Micros, Austria) was used. The structure of healthy *Chlorella* sp. is spherical (2–5 µm); however, the structure and size of *Chlorella* sp. can change due to exposure to toxic substances such as NPs (Khoshnamvand et al. 2024). As can be seen in the control (Fig. 4), micro-algae are round and separate, having a good density, and their sizes seem appropriate. *C. vulgaris* exposed to TBBPA (6 mg/L) showed a decrease in cell density and started losing pigments. The culture started to decline and the medium's color reduced by the second day following the addition of TBBPA. From the third day onwards, growth was observed only at 6 mg/L of TBBPA. However, at higher concentrations (10 mg/L), the culture color diminished by the second day post-inoculation in all culture flasks containing TBBPA. Under the light microscope (Micros-MCX300LED, Austria), the cell morphology (*Chlorella* Sp.) in the control and 2 mg/L flask remained unchanged. At a higher concentration of 10 mg/L TBBPA, the culture medium was totally faded, the cells were dead and ruptured, and no chlorophyll content was detected in all the high concentrations (8mg and 10mg) of TBBPA. (Fig 4). Similarly, *C. vulgaris* has greater sensitivity to fludioxonil than *Scenedesmus acutus*, *Scenedesmus obliquus*, *Lemna minor*, and other aquatic organisms (Verdisson et al., 2001; Dewez et al., 2005). In comparison to other pollutants, including spherical mancozeb, chlorothalonil, and pyrimethanil, fludioxonil exhibits a more potent growth inhibiting effect on *C. vulgaris* (Xi et al., 2019). The inhibitory effects on the growth of *C. vulgaris* linked to elevated TBBPA concentrations in aquatic ecosystems necessitate additional investigation.

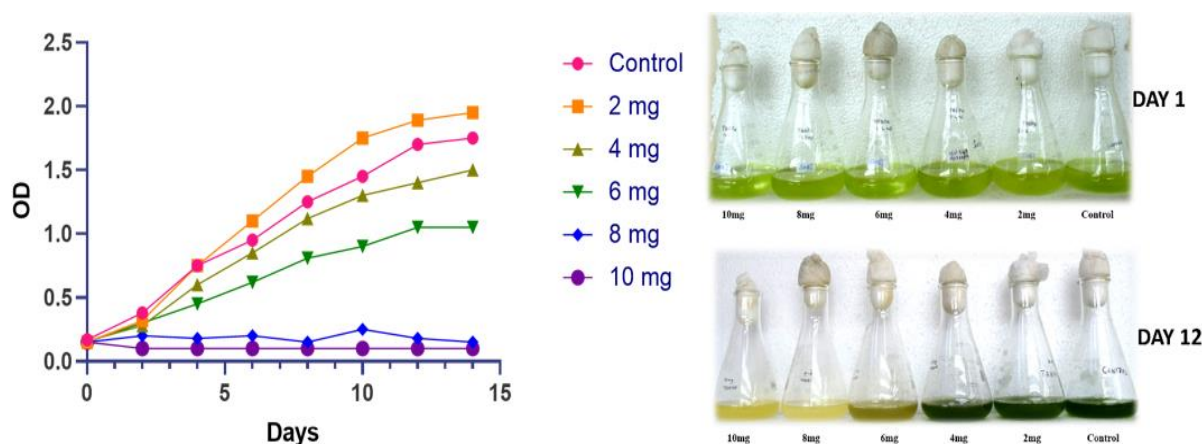


Fig. 3. Dose–response growth curve of the *Chlorella* sp. to TBBPA with different concentration

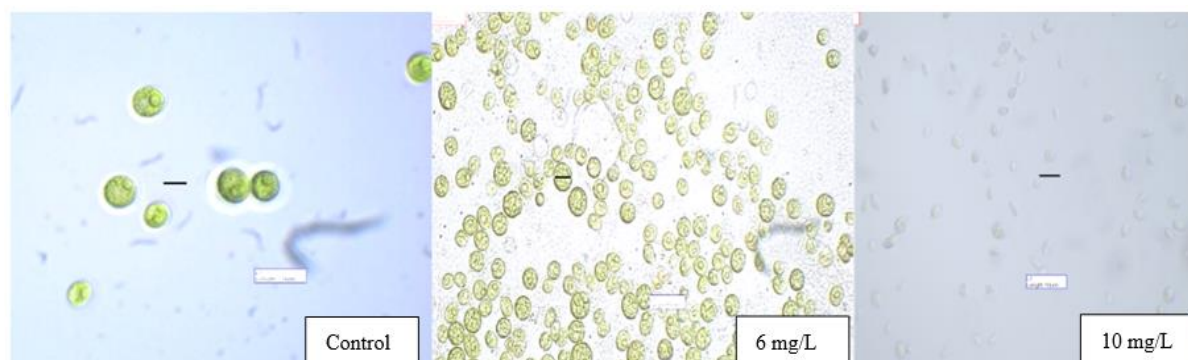


Fig. 4. Cell morphology of *Chlorella* sp. After exposed to TBBPA with different concentrations

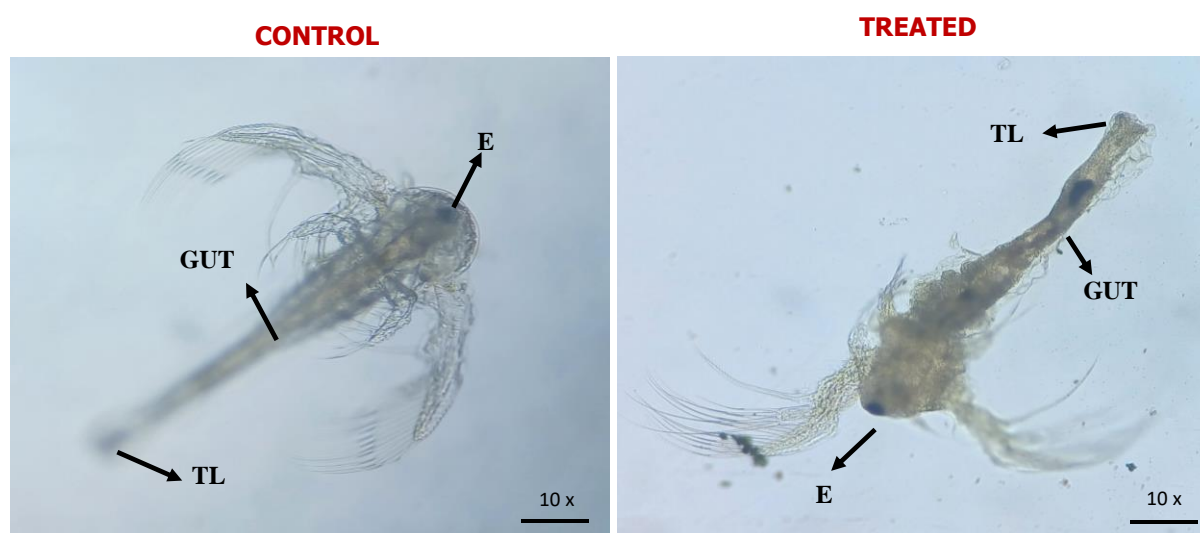


Fig. 5. Toxicity bioassay of TBBPA on *Artemia nauplii* (*A. salina*)

Table 1. Non-toxicity bioassay of TBBPA on *Artemia salina* (Brine shrimp)

Animal name	Sample	LC ₅₀ µg/mL (LCL – UCL)	LC ₉₀ µg/mL (LCL – UCL)	χ ²
<i>Artemia salina</i>	TBBPA	128.21 107.86 – 182.17	189.58 150.43 – 299.35	2.907

LCL lower confidence limit, UCL upper confidence limit, χ² Chi square test, p<0.05, level of significance, values are mean ± SD of three replicates

3.4 Toxicity Assay

The toxicity assay on *Artemia salina* demonstrated moderate mortality rates of 58% after 24 hours. No structural damage was detected in the nauplii. The mortality of brine shrimp after a 24-hour treatment yielded LC₅₀ and LC₉₀ values of 128.21 µg/mL (range: 107.86–182.17 µg/mL) and 189.58 µg/mL (range: 150.43–299.35 µg/mL) (Table 1).

The morphological effects of TBBPA on *Artemia salina* were photo documented using light microscope (Micros, Austria) and are presented in Fig. 5. Differences between the control and treated were observed. In control group *artemia salina* showed morphologically healthy body structure, including clearly visible eyes(E), straight and intact tail region (TL) and a clear undistorted gut (GUT). In the treated group, TBBPA caused morphological abnormalities in the species. The nauplii showed deformation in the tail region and swelling. The Gut region

appeared possible damage to digestive processes. This structural deformity indicates that TBBPA showed a toxic impact on the *A. salina* early development, which affected the physiological and digestive function.

4. CONCLUSION

This study looked at how TBBPA affected the growth of *Chlorella* sp. and *Artemia salina*, showing different levels of growth slowdown in both during the exposure. We also observed that TBBPA affected the shape of *Chlorella* sp. and *Artemia salina*, leading to a 58% death rate within 24 hours. In aquatic environments, *Chlorella* sp. and *A. salina* are important primary producers that are utilized as an energy source, while TBBPA, which is commonly found, raises concerns about its toxicity to the environment. Consequently, it is essential to examine the harmful impacts of TBBPA on aquatic organisms. The results show that TBBPA is a serious environmental threat to water ecosystems, emphasising the need for stricter rules and more studies on the ecological risks of brominated flame retardants.

DISCLAIMER (ARTIFICIAL INTELLIGENCE)

Author(s) hereby declares that NO generative AI technologies such as Large Language Models (ChatGPT, COPILOT, etc) and text-to-image generators have been used during writing or editing of this manuscript.

DATA AVAILABILITY STATEMENT

The data that support the findings of the study are available from the corresponding author upon responsible request.

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COMPETING INTERESTS

Authors have declared that no competing interests exist.

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